Status of the Northern Myotis (Myotis septentrionalis) in Alberta:
Update 2009

Alberta Wildlife Status Report No. 3 (Update 2009)
Status of the Northern Myotis
(Myotis septentrionalis) in Alberta:

Update 2009

Prepared for:
Alberta Sustainable Resource Development (SRD)
Alberta Conservation Association (ACA)

Update prepared by:
Cori Lausen

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PREFACE

Every five years, the Fish and Wildlife Division of Alberta Sustainable Resource Development reviews the general status of wildlife species in Alberta. These overviews, which have been conducted in 1991 (The Status of Alberta Wildlife), 1996 (The Status of Alberta Wildlife), 2000 (The General Status of Alberta Wild Species 2000), and 2005 (The General Status of Alberta Wild Species 2005) assign individual species “ranks” that reflect the perceived level of risk to populations that occur in the province. Such designations are determined from extensive consultations with professional and amateur biologists, and from a variety of readily available sources of population data. A key objective of these reviews is to identify species that may be considered for more detailed status determinations.

The Alberta Wildlife Status Report Series is an extension of the general status exercise, and provides comprehensive current summaries of the biological status of selected wildlife species in Alberta. Priority is given to species that are At Risk or May Be At Risk in the province, that are of uncertain status (Undetermined), or that are considered to be at risk at a national level by the Committee on the Status of Endangered Wildlife in Canada (COSEWIC).

Reports in this series are published and distributed by the Alberta Conservation Association and the Fish and Wildlife Division of Alberta Sustainable Resource Development. They are intended to provide detailed and up-to-date information that will be useful to resource professionals for managing populations of species and their habitats in the province. The reports are also designed to provide current information that will assist Alberta’s Endangered Species Conservation Committee in identifying species that may be formally designated as Endangered or Threatened under Alberta’s Wildlife Act. To achieve these goals, the reports have been authored and/or reviewed by individuals with unique local expertise in the biology and management of each species.
EXECUTIVE SUMMARY

The northern myotis (*Myotis septentrionalis*; previously known as the northern long-eared bat) is a forest-dependent bat commonly found in eastern Canada and the United States. It is encountered less frequently in western Canada, including Alberta. At present, the northern myotis is considered *May Be At Risk* in Alberta. This general status designation was based on the relative rarity of this species in the province, and its apparent reliance on mature forest habitats, which are affected by current forestry practices. However, recent captures in Alberta’s boreal forest bring into question whether this species may be relatively abundant in some areas of northern Alberta; further standardized survey effort across central and northern Alberta is needed to establish relative abundance, and determine if this species is widespread across its range or whether it has a patchy distribution.

The northern myotis has a broad distribution in Alberta, north of a V-shaped line extending from Wainwright to the northernmost reaches of the Red Deer River and west to Jasper. Within this area, the species is reported most often from the Boreal Forest and northern portion of the Foothills natural regions, although there are several records from the Parkland Natural Region. During the winter, the northern myotis hibernates with other bat species in caves or abandoned mines. Two hibernacula have been found in Alberta: one in a cave in Wood Buffalo National Park, and one in a cave near Cadomin.

Consistent among studies (all conducted outside of Alberta) is the tendency of this species, especially females, to roost in cracks, hollows, or under bark of deciduous trees in deciduous-dominated stands; however, roosts in conifer trees are also reported, and in eastern Canada, male northern myotis tend to select roosts in coniferous trees in conifer-dominated stands. Early studies from other parts of the species’ range suggested that roost trees tended to be taller and wider than what was randomly available, and that these trees were generally found in old-growth forest stands. More recent studies from other parts of its range contradict some of these findings, describing the northern myotis as more of a generalist, using roost trees of various heights and diameters, and even using remnant snags within young (under 5 years) stands. The apparent discrepancies among roosting and foraging studies suggest that either the northern myotis may be more adaptable than previously believed, or that there are extensive local variations in behaviour and ecology across this species’ range, thereby necessitating study of the habitat requirements of the northern myotis in Alberta.

Factors that may limit northern myotis populations include availability and security of summer roosting and foraging habitat and of winter hibernacula, and climate change. Until summer and winter habitat requirements are known for this species in Alberta, it is unclear whether current forestry and northern development practices affect long-term population viability. Climate change is a concern to northern myotis in Alberta because of the prediction that mean tree age, and consequently tree size, in undisturbed forests will continue to decline as a result of recently documented increasing tree mortality in western North America. Another uncertainty for northern myotis is the potential spread of White Nose Syndrome into Alberta; this fungal infection causes high rates of bat mortality in hibernacula.
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# TABLE OF CONTENTS

PREFACE ........................................................................................................................................................................ iii  
EXECUTIVE SUMMARY ......................................................................................................................................................... iv  
ACKNOWLEDGEMENTS ............................................................................................................................................................ v  
INTRODUCTION ....................................................................................................................................................................... 1  
HABITAT .................................................................................................................................................................................. 1  
  1. Winter Hibernacula .......................................................................................................................................................... 1  
  2. Summer Roosting and Foraging Habitat ...................................................................................................................... 2  
CONSERVATION BIOLOGY ....................................................................................................................................................... 4  
  1. Identification ............................................................................................................................................................... 4  
  2. Hibernation ................................................................................................................................................................. 5  
  3. Reproduction and Survival .............................................................................................................................................. 6  
  4. Foraging ........................................................................................................................................................................ 6  
DISTRIBUTION ......................................................................................................................................................................... 7  
  1. Alberta ............................................................................................................................................................................ 7  
  2. Other Areas ................................................................................................................................................................. 7  
POPULATION SIZE AND TRENDS ........................................................................................................................................... 10  
  1. Challenges ............................................................................................................................................................... 10  
  2. Population Size .......................................................................................................................................................... 10  
  3. Population Trends ...................................................................................................................................................... 11  
LIMITING FACTORS ............................................................................................................................................................... 13  
  1. Availability of Summer Roosting Habitat ..................................................................................................................... 13  
  2. Hibernacula ............................................................................................................................................................... 15  
  3. White Nose Syndrome ................................................................................................................................................ 17  
  4. Climate Change .......................................................................................................................................................... 17
TABLE OF CONTENTS cont.

STATUS DESIGNATIONS ........................................................................................................18
  1. Alberta .......................................................................................................................18
  2. Other Areas ...............................................................................................................18

RECENT MANAGEMENT IN ALBERTA ..............................................................................19

SYNTHESIS .......................................................................................................................19

LITERATURE CITED .........................................................................................................22

APPENDIX 1 Definitions of status ranks and legal designations .......................................31

APPENDIX 2 Site locations of the northern myotis in Alberta .......................................33

LIST OF FIGURES

Figure 1 Site records and presumed distribution of the northern myotis in Alberta ..........8

Figure 2 Range of the northern myotis in North America ...............................................9

LIST OF TABLES

Table 1 Summary by region of published studies reporting capture of the northern myotis in Canada .........................................................................................................12
INTRODUCTION

The northern myotis (Myotis septentrionalis; previously known as northern long-eared bat, Wilson and Reeder 2005) is found in many regions of Canada. Although there are numerous records of its presence in eastern Canada and the United States (Barbour and Davis 1969, Dobbyn 1994, Kurta 1995, van Zyll de Jong 1985), it has been recorded more sporadically in western Canada (Barbour and Davis 1969, Nagorsen and Brigham 1993, Smith 1993). Presently, the northern myotis is considered May Be At Risk* in Alberta (Alberta Sustainable Resource Development 2007).

This report, an update of the original status report (Caceres and Pybus 1997), summarizes information relevant to the status of the northern myotis in Alberta. Before proceeding, it should be noted that prior to 1980, the northern myotis was considered to be a subspecies of M. keenii and was occasionally referred to as “Keen’s bat,” or the “eastern long-eared bat” (Fitch and Shump 1979, van Zyll de Jong 1979). Presently, M. keenii is restricted to the Pacific Northwest and is now known to be genetically distinct from the northern myotis (Dewey 2006); any references to this species from outside that area refer to M. septentrionalis (van Zyll de Jong 1979).

It should also be noted that since the original status report for this species (Caceres and Pybus 1997) was published, acoustic and mistnetting surveys of bats in northeastern Alberta have increased, coinciding with the increased industrial development in that region. No habitat studies have been done on this species in Alberta; therefore, several roost selection studies done in other parts of its range provide our only foundation upon which to understand this species’ habitat in Alberta.

* See Appendix 1 for definitions of selected status designations.

HABITAT

Habitat selection by the northern myotis can be divided into two major components: 1) winter hibernacula; and 2) summer roosting and foraging habitat.

1. Winter Hibernacula. - The northern myotis hibernates in caves or abandoned mines (Caire et al. 1979, Griffin 1940, Hitchcock 1949, Mills 1971, Schowalter 1980, Swanson and Evans 1936, Thomas 1995). Swarming and mating are thought to occur in the vicinity of hibernacula prior to hibernation (Caceres and Barclay 2000). Within a cave, northern myotis generally occurs with other species of bats, but is thought to form a small proportion of the total hibernating population (Griffin 1940, Hitchcock 1949, Schowalter 1979, Thomas 1993). The largest hibernating group of northern myotis in Canada was approximately 300 individuals in a cave in Quebec (Thomas 1993); in this hibernaculum, northern myotis made up approximately 25% of the total number of hibernating bats. Northern myotis may be overlooked in a hibernaculum because this species tends to roost in deep, hard-to-sample crevices (reviewed in Caceres and Barclay 2000). It is, however, unclear whether this roosting behaviour would result in gross under-estimates of the relative abundance of northern myotis in hibernacula.

There are two known hibernacula used by northern myotis in Alberta: Cadomin Cave (53°00’N; 117°20’W; Schowalter 1980), and Wood Buffalo National Park (60°00’N; 114°00’W; Schowalter 1979). The hibernacula were shared with little brown myotis (M. lucifugus) and long-legged myotis (M. volans), but specific site preferences within the hibernacula differed among species. Northern myotis made up zero to three percent of the total number of bats collected at any one time at these sites. In Alberta, bats often reach the hibernacula in late August and early September (Schowalter 1980).
In general, bats choose hibernaculum sites that provide a relatively constant, low (0°C to 4°C) temperature with high humidity and no air currents. Within these sites, northern myotis usually hang singly in small narrow crevices (Caire et al. 1979, Kurta 1995). Anecdotal evidence suggests that northern myotis prefer moister and cooler sites (caves/mines or regions of a cave/mine) than little brown myotis in the same hibernaculum (Barbour and Davis 1969, Hitchcock 1949).

2. Summer Roosting and Foraging Habitat.

Across its range, northern myotis uses a variety of tree species for roosting (Owen et al. 2002), suggesting it is not the tree species but rather the physical and microclimatic characteristics that dictate roost selection (Chung-MacCoubrey 2005). Mature forest stands provide important roosting habitat for this species (Vonhof and Wilkinson 1999); however, the exact tree characteristics selected by this species are unclear. Characteristics of trees used as roost sites in Alberta have not been investigated, and studies elsewhere have produced mixed results regarding roost tree selection. In general, this species is dependent on forests, foraging and commuting in the forest interior (Broders et al. 2006, Carter and Feldhamer 2005, Jung et al. 1999, Lacki and Hutchinson 1999, Owen et al. 2003, Patriquin and Barclay 2003), but also commuting along edge habitats, rarely (if ever) flying through open non-forested areas (Henderson 2007, Hogberg et al. 2002).

Roosting occurs only in intact forest (i.e., forest interior), and unlike most other species of tree-roosting bat, northern myotis select roost trees with a higher degree of canopy cover (stem density, vegetative clutter) than around potential roost trees (Foster and Kurta 1999, Menzel et al. 2002, Owen et al. 2002). This is likely because of their great agility (reviewed in Caceres and Barclay 2000). This species’ ability to glean insects from vegetation, in addition to aerial hawking of prey, makes them well adapted for forest interior foraging (Ratcliffe and Dawson 2003), and has likely resulted in their apparent dependency on intact forests. Henderson and Broders (2008) determined that northern myotis do not fly more than 78 m from an edge of intact forest, thereby avoiding open, unforested areas. In general, cavity-roosting bats select roost trees based on proximity to water (Kalcounis-Rüppell et al. 2005). In Illinois, during the driest periods, roosts of northern myotis were within 750 m of a source of fresh water (Carter and Feldhamer 2005).

Northern myotis have been found roosting in crevices behind peeling bark, cavities in partially decayed trees, or cracks in trees (Carter and Feldhamer 2005, Clark et al. 1987, Foster and Kurta 1999, Menzel et al. 2002, Mumford and Cope 1964, Vonhof et al. 1997). While some studies report use of mainly dead trees (Broders and Forbes 2004, Menzel et al. 2002, Sasse and Pekins 1996), others report equal use of dead and living trees (Carter and Feldhamer 2005, Foster and Kurta 1999). Roosts in mid-decay stages have frequently been reported (Broders and Forbes 2004, Lacki and Schwierjohann 2001). Northern myotis has been shown repeatedly to select roosting areas containing a high density of snags (Owen et al. 2002); this is especially true for females roosting in colonies (Lacki and Schwierjohann 2001). Solitary-roosting individuals show greater flexibility, roosting in living trees more often than colonial-roosting individuals (Lacki and Schwierjohann 2001).

Early roosting studies of northern myotis found that tall, wide-diameter, partially dead trees with a high percentage of the bark remaining are favoured over other seemingly suitable trees (Caceres 1998, Sasse and Pekins 1996), corroborating studies of other tree-roosting bats (Kalcounis-Rüppell et al. 2005, Vonhof and Barclay 1996). However, more recent roosting studies on this species report that researchers are unable to discriminate between roost trees and potential roost trees (those with at least one cavity and/or of minimum diameter at breast-
height [dbh]; Foster and Kurta 1999, Vonhof et al. 1997). Roost trees are not always tall, with roosts as low as 1 m from the ground (Cryan et al. 2001). Unfortunately, all of these roosting studies have occurred in areas substantially south of northern Alberta, where roosting conditions are inevitably different because of climate. It is therefore difficult to extrapolate any of these results to the roost preferences of northern myotis in northern Alberta.

A sex-specific roosting study in eastern Canada (Broders and Forbes 2004) documented males roosting in smaller trees (mean dbh 32.0±1.6SE), whereas females typically roosted in cavities of trees with larger dbh (mean dbh 43.8±1.8SE). In general, it has been determined that females tend to roost in groups in mature, shade-tolerant deciduous stands (Broders and Forbes 2004; Foster and Kurta 1999; Menzel et al. 2002; Sasse and Pekins 1996), whereas males typically roost solitarily in either deciduous or coniferous trees (Broders and Forbes 2004; Ford et al. 2006; Jung et al. 2004; Lacki and Schwierjohann 2001). Although coniferous trees appear to be important to male northern myotis, conifer trees are rarely selected as roosts by females (Broders and Forbes 2004, Henderson et al. 2008).

Warm roosts are particularly important to reproductive females and juveniles. It is likely that adult female northern myotis, like other bats, use torpor (body cooling) to a lesser extent during reproduction in order to maximize fetal development and milk production (Racey 1973, Wilde et al. 1999). Newborn Myotis bats grow rapidly and generally begin flying within a month after their birth (Fenton 1983). Maximum energy is directed toward growth, and a maternity roost that maintains a warm microclimate will maximize juvenile growth and development (Humphrey et al. 1977). Roosts may be selected based on microclimate to meet thermoregulatory requirements during reproduction (e.g., Lausen and Barclay 2003). Given that northern myotis roosts tend to be lower in the canopy, and in more cluttered (i.e., dense or structurally complex vegetation) and thus more shaded sites than are typical for many other tree-roosting species, it has been suggested that solar radiation may be a less important roost criterion than it is for other bat species (Menzel et al. 2002). Unfortunately, no thermoregulatory studies have been done on this species to know whether their patterns of torpor and roost microclimate selection differ from other species. In the Greater Fundy area of New Brunswick, northern myotis consistently give birth several weeks later than building-roosting little brown myotis in the area (Broders et al. 2006). This may reflect roost microclimate, and suggests differences in selection pressure and perhaps reproductive success between the two species (Broders et al. 2006, Lausen and Barclay 2006b).

Individual northern myotis move between a number of roost trees, switching every two days on average (Foster and Kurta 1999), although they seem to switch less often during lactation (every 5 days on average; Menzel et al. 2002). Roosts can be within a few hundred metres of each other (Crampton 1995, Vonhof and Barclay 1996), but consecutive roosts up to 2 km apart have been reported (Foster and Kurta 1999). Females have been found to have roosting and foraging ranges that are 6.1 and 3.4 times larger, respectively, than those of males (Broders et al. 2006).

Although earlier capture records suggested this species was dependent on mature or old-growth forest stands (as summarized in Caceres and Pybus 1997), more recent widespread capture efforts in Alberta and elsewhere have found northern myotis also in young stands and disturbed forest regimes (Crampton and Barclay 1998, Cryan et al. 2001, Foster and Kurta 1999, Henderson and Broders 2008, Henderson et al. 2008, Lausen et al. 2006a, Menzel et al. 2002). In disturbed landscapes, such as an agriculture-forest mosaic, this species has been found in remnants of intact forest,
provided minimum woodlot sizes are retained. Minimum forest fragment sizes for Prince Edward Island have been established (see Limiting Factors – Summer Roosts for further details; Henderson et al. 2008), but no such study has taken place in western Canada. In a fragmented forested landscape, foraging range sizes for female northern myotis are substantially smaller (9% – 13% of the foraging ranges of northern bats in intact forest; Broders et al. 2006, Henderson 2007, Owen et al. 2003). It is likely that population density decreases in patchy landscapes in response to limited roosts and foraging opportunities; however, population density is difficult to measure. In Alberta, smaller intact forest fragments have been associated with less bat activity than have larger blocks of forest retention (Marinelli 2000). Although reduced mobility and home range sizes have been studied in fragmented forest environments, no attempt has been made to understand population size and changes in bat density in relation to these landscape disturbances.

Presumably, this species has evolved with a natural disturbance regime. Historically, fire has created a mosaic of forest habitat patches, leaving live and dead trees in clumped patches (Hansson 1992, EMEND 2006). After small patches of forest burn, local shifts in bat foraging and roosting behaviour are likely to occur in response to available habitat (Patriquin and Barclay 2003).

Reproductive female northern myotis may gather in various-sized colonies, showing some degree of social organization (evidence of non-random sub-groupings of the colony; Garroway and Broders 2007). Maternal colonies may be in buildings (Barbour and Davis 1969, Brandon 1961, Henderson and Broders 2008), or in cavities in trees; mean colony size in Michigan was reported to be 17±2 bats (n = 65 observations; Foster and Kurta 1999), with colonies as large as 88 in West Virginia (Owen et al. 2002). Most males or non-reproductive females roost singly or in small (<10) groups in trees; however, they may also roost in or on buildings, or in caves or other available crevices (reviewed in Caceres and Barclay 2000). The relative importance of human-made structures for day roosts for northern myotis in Alberta is unknown. The only putative building-roosting colonies of this species were reported by Vujnovic (2001); however, genetic evidence post-reporting has since revealed an error in species identification; the building-roosting northern myotis reported in Vujnovic (2001) are in fact little brown myotis (D. Vujnovic, pers. comm.). Genetic species confirmation together with thorough morphological quantification of building-roosting Myotis from central and northern Alberta is encouraged to better document building-roosting colonies of bats in the province.

**CONSERVATION BIOLOGY**

1. **Identification.** - The northern myotis is one of five species of *Myotis* that occur in Alberta. The species is most likely to be confused with the little brown myotis or the long-eared myotis (*M. evotis*), but can sometimes be distinguished from these species on the basis of ear length (14 mm - 19 mm versus 9 mm - 17 mm and 17 mm - 22 mm in the little brown myotis and the long-eared myotis, respectively; Nagorsen and Brigham 1993). Unfortunately, ear length is a difficult measurement to make correctly (ruler placed along the lateral edge inside the pinna and carefully but firmly rooted in the basal part of the ear), and even measurements made on the same animal by different observers can differ by several mm (D. Nagorsen and C. Lausen, unpubl. data). Even with accurate pinna measurements, ear length overlap among the species can preclude conclusive species identification based on this morphological trait alone. In northern Alberta, tragus (small pointed projection in the middle of the external ear) shape, absence of dark face-mask, subtle membrane/fur colour contrast, and presence of a keeled calcar (spur of cartilage arising...
from the inside of the ankle) can assist in identifying northern myotis. Although little brown myotis can have long ears (Nagorsen and Brigham 1993), the tragus is generally blunt, wide and often lobular in shape. This is in contrast to northern myotis, which have a narrow triangular-shaped tragus that ends in a relatively sharp point (D. Nagorsen and C. Lausen, unpubl. data). Additionally, northern myotis tend not to have obvious masks across their eyes and their brown fur colour does not sharply contrast with their brown membranes; this differs from the long-eared myotis, which have obvious masks and dark brown pinnae and membranes (pinnae of long-eared myotis tend to be very glossy), contrasting sharply against their pale pelage (Nagorsen and Brigham 1993, van Zyll de Jong 1985). Finally, although not always present, a very small keel (triangular flap of tissue) can sometimes be found on one or both calcars of northern myotis, aiding in their differentiation from little brown myotis (Lausen et al. 2008).

Northern myotis, although morphologically similar to other *Myotis* in Alberta, are genetically distinct, with mitochondrial DNA sequences that show high levels of divergence from other Alberta *Myotis* (Dewey 2006). As such, the only definitive way to differentiate this bat from other similar species is through genetic analysis.

It has been hypothesized that northern myotis have been misidentified as little brown myotis for many years in northern Alberta (K. Patriquin, pers. comm.). Genetic sampling is now allowing for conclusive identification of northern myotis; Grindal et al. (submitted) found that 86% and 83% of recent field identifications of little brown myotis and northern myotis, respectively, by experienced bat biologists were correctly identified in the field when verified by genetic testing. This confirms that error in field identification has taken place, and suggests that biologists less experienced with bat identification may have made substantial misidentifications of these species in past surveys.

The average size reported for the northern myotis is 7.4 g with forearm length of 36.4 mm (van Zyll de Jong 1985). Females tend to be slightly larger than males. In addition, a full-term fetus may add up to 25% to the body mass of a pregnant female. However, size is a highly variable characteristic and adult northern myotis as small as 5 g, with a 34-mm forearm, have been caught (Nagorsen and Brigham 1993, M. C. Caceres, unpubl. data).

It is difficult to differentiate little brown myotis, northern myotis, and long-eared myotis in flight, because their echolocation calls broadly overlap at approximately 40 kHz (Fenton et al. 1983). Although there are some features of the call sonogram that can allow the species to be differentiated in some cases, overlap of call characteristics will result in an under- or over-estimate of the species, depending on the rule-set employed (C. Lausen, unpubl. data). This is most relevant when analyzing zero-crossing data (sonograms as recorded by the Australian-made Anabat detector [Titley Electronics]), given that sound amplitude is not recorded (amplitude is recorded with other types of ultrasonic detectors such as Pettersson time expansion units [Pettersson Electronik, Sweden], although these units are currently not well suited for passive monitoring).

2. Hibernation. - Female *Myotis* bats begin visiting their hibernaculum between August and early October. Suitable hibernaculum sites have a narrow range of environmental conditions and thus bats may have to travel long distances to reach an appropriate site. Although the distance travelled between the summer habitat and the hibernaculum site is generally not known, banding recoveries indicate that some little brown myotis that summer in north-central Alberta travel to the Cadomin area to hibernate (Fish and Wildlife Division, unpubl. data). It seems reasonable to assume that
northern myotis also may travel long distances to reach their hibernacula; the longest distance reported for this species between summer and hibernating areas in B.C. is 56 km (Nagorsen and Brigham 1993).

Male and female *Myotis* spp. return to the same hibernaculum year after year. Some little brown myotis banded in Cadomin Cave in the late 1970s (Schowalter 1980) have been recaptured or resighted as recently as 2009 (D. Hobson, pers. comm.).

Swarming and mating in *Myotis* spp. occur in the fall, generally at the hibernaculum site. Females then store sperm over the winter, and a single egg is fertilized when the female auras from torpor and leaves the hibernaculum in the spring (Racey 1982). Gestation (fetal development) is approximately 60 days and birth occurs in late June and July (van Zyll de Jong 1985).

During summer, bats must accumulate enough resources to maintain the high metabolic costs associated with flight and reproduction; in late summer and fall, they feed to gain sufficient fat reserves for winter hibernation. Female bats must accumulate greater energy reserves than males in order to produce and care for the offspring (Barclay 1991). Lactating females need adequate food to meet their own energy needs as well as to produce milk for the pup. Canadian species of *Myotis* generally add 30% to 40% of their summer mass in fat to use as energy reserves during hibernation (Nagorsen and Brigham 1993, Thomas et al. 1990).

### 3. Reproduction and Survival.

Unlike other small mammals, all North American *Myotis* spp., including the northern myotis, are relatively long-lived. The record age of a northern myotis in the wild is 18.5 years (Hall et al. 1957). Also in contrast to most small mammals, each female bat produces only one offspring each year. Although anecdotal records of predation on bats have not included the northern myotis, it is likely that owls, magpies, and snakes prey upon them occasionally, as these are predators on other bats in Alberta.

Environmental conditions influence the ability of a female bat to successfully raise an offspring in any given season. When conditions such as high rainfall, cold weather, or windy evenings reduce insect availability, females may not be able to gather enough food; this can delay growth of offspring, and if it occurs in spring, it may cause females to delay or forego reproduction entirely for that season (Grindal et al. 1992). Most mortality in North American bat species occurs in the juvenile age class and many pups do not survive their first year (Tuttle and Stevenson 1982). Additional mortality occurs during the hibernation period. If winters are harsh, or if bats cannot obtain the fat reserves necessary to survive this period, they may starve. Juveniles are particularly vulnerable during hibernation because they have a relatively short time after their birth to build up fat reserves (Nagorsen and Brigham 1993). Especially in northern climates where the growing season is short, delayed parturition could have a negative impact on pups’ first-year survival; little is known about reproductive patterns and survival rates in the northern distribution of most bats. Reproductive data are not housed in the provincial database (Alberta Fish and Wildlife Management Information System), therefore, the age of first reproduction and reproductive rates of northern myotis females in Alberta are unknown. Other temperate bat species generally reproduce when they are one or two years of age (Altringham 1996).

### 4. Foraging.

Like all bat species in Canada, the northern myotis feeds exclusively on insects. *Myotis* bats are generalist insectivores, and their diet is limited by the size, behaviour, and quality (e.g., carapace hardness) of the insects they are capable of catching. Although moths and beetles make up the majority of the diet of *Myotis* spp., mosquitoes, black flies, and other species considered to be pests also
are consumed. Generally, northern myotis uses various food items, including species of Hemiptera (true bugs), Lepidoptera (moths), Hymenoptera (bees and wasps), Diptera (flies), and Homoptera (includes leafhoppers and aphids; van Zyll de Jong 1985).

Within the group of Myotis bats, differences in foraging strategy (Fenton 1990, Fenton and Bell 1979) have evolved to reduce overlap and competition. Most insectivorous bats feed on flying insects (these bats are called “hawkers”). Other bats (known as “gleaners”) use passive listening as well as echolocation to locate and capture insects resting on leaves, tree trunks or against buildings. This technique is an apparent foraging advantage in cooler conditions when the abundance of flying insects is reduced (Taylor 1963). Northern myotis are opportunistic and will hawk insects from the air or glean prey from substrates (Ratcliffe and Dawson 2003).

**DISTRIBUTION**

1. **Alberta.** - In Alberta, the northern myotis is most often reported in the Boreal Forest Natural Region, and the northern section of the Foothills Natural Region (Figure 1, Appendix 2; Alberta Natural Heritage Information Centre 2006). Northern myotis have been captured at Cadomin Cave (Schowalter 1979, H. Smith, pers. comm.), and 16 of 30 northern myotis specimens in the Provincial Museum of Alberta were collected at this site. There also are limited records from the Peace River Parkland and Central Parkland subregions. Most recently, a large number of captures have been made in the oil sands area near Fort McMurray (Figure 1, Appendix 2).

It is not clear from the capture data for Alberta whether northern myotis is evenly distributed across the northern part of the province, but it is likely scattered across the Boreal Forest Natural Region, and at least some of the Parkland and Foothills natural regions (Figure 1). Given that bats tend to be selective in the micro-habitats they occupy (e.g., particular roost conditions are selected within forest stands), and that the Boreal region is not uniformly composed of suitable roost trees, local distributions are probably patchy. Therefore, local concentrations of northern myotis, such as those documented during extensive sampling in the oil sands area near Fort McMurray (S. Grindal, pers. comm.) are likely to reflect pockets of suitable habitat; this relative abundance of northern myotis may not be representative of the entire Alberta distribution of northern myotis. As expected, based on percent forest cover, this species is far less common in the Parkland region compared to the Boreal region.

The area of occupancy of northern myotis is 132 km². This reflects the area summation of 2 km x 2 km squares that have had mistnet sampling within them, with at least one northern myotis being captured (or in the case of museum records, at least one voucher being taken). The extent of occurrence is 357 026 km², representing the area of the polygon encompassing all known records. For this latter calculation, potentially unoccupied and unsuitable habitats may be included.

2. **Other Areas.** - The northern myotis is found across the southern forested regions of Canada (except in B.C.; Figure 2), but is captured more frequently in eastern Canada and the United States than in western Canada. Records of this species in western regions are on the rise and may reflect increased survey efforts and/or an increased ability to distinguish northern myotis from other similar Myotis. As more widespread inventory efforts continue, and genetic identification of captures becomes more common, accurate distribution maps and relative abundance estimates will follow.

The species has been found regularly in southern Ontario, southern Quebec, and in the Maritime provinces, as well as in eastern regions of the United States (Dobbyn 1994, Godin 1977,
Figure 1: Site records and presumed distribution of the northern myotis in Alberta. Numbered sites are described in Appendix 2.
Figure 2: Range of the northern myotis in North America (adapted from Caceres and Barclay 2000).
The range of the northern myotis in the United States extends from the Canadian border south to northern Florida (Rice 1955) and west into North Dakota, South Dakota, Nebraska, and Wyoming (Knox Jones et al. 1983, Turner 1974). In Canada, the documented range of the northern myotis extends north into central Quebec, and northern parts of Ontario, Manitoba, Saskatchewan, and Alberta, as well as the southwest corner of the Northwest Territories and the southeast corner of the Yukon (Jung et al. 2006, Lausen 2006b, Lausen et al. 2008, Nagorsen and Brigham 1993, Nagorsen and Nash 1984, Smith 1993, van Zyll de Jong 1985, Wrigley 1974). Currently the western boundary of the range appears to be the Coastal Mountains in northern British Columbia, and the west slope of the Rocky Mountains in more southern reaches of B.C. (near Revelstoke, B.C.; van Zyll de Jong et al. 1980, B.C. Ministry of Environment 2008).

Overall, the known range of the northern myotis is primarily restricted to forested regions, with a distinct avoidance of the prairies, coastal rainforests, and southern Rockies. Occasionally, northern myotis is captured in the central plains of the United States, but these situations occur in forest remnants or in caves used for hibernation (Turner 1974). Henderson et al. (2008) reported loss of forest cover and specifically the loss of deciduous stands to be limiting the distribution of northern myotis in eastern Canada; on Prince Edward Island the northern myotis is restricted to remnant fragments of forest of sufficient size with sufficient tree cover and percentage of deciduous trees.

**POPULATION SIZE AND TRENDS**

1. **Challenges.** - The life history of bats (i.e., low reproductive rates, high overwinter juvenile mortality; Barclay et al. 2004) makes their populations vulnerable to declines, and therefore being able to monitor trends in populations of bats is highly desirable (O’Shea and Bogan 2003). Estimates of population size of northern myotis are hampered by the fact that the species is a forest-dwelling bat that is difficult to detect. Because this bat tends to forage in the forest interior, mistnet surveys must target cluttered areas; northern myotis are less likely to be captured over water than are little brown myotis, requiring different mistnetting strategies than are used to catch other myotis (Broders et al. 2006, Jung et al. 1999). Similarly, acoustic surveys must place detectors on both forest edges and in the forest interior rather than in clearings near open water bodies to increase the chances of recording northern myotis. Even then, acoustic recordings will usually under-represent this species (Broders et al. 2003) because of the low intensity ultrasound produced by this bat compared to other species (Faure et al. 1993). Although some species of bats in northern Alberta are detected and recorded by bat detectors up to 40 m away from a bat, northern myotis generally must approach within a few metres of the detector microphone to be recorded (C. Lausen, pers. obs., D. Nagorsen, pers. comm.). In northern Alberta, echolocation calls of northern myotis overlap with those of little brown myotis (Broders et al. 2004), further limiting the ability to gather information about this species from acoustic recordings.

2. **Population Size.** - Although capture numbers for northern myotis have been increasing over recent years as a result of increased bat inventory efforts in the north, especially in the northeast in conjunction with industrial developments, it is impossible to estimate the provincial population size of the northern myotis. Survey efforts in the oil sands area near Fort McMurray have led some biologists doing work in this area to suggest this species, which is abundant in relation to other northern bat species in this region, is relatively more abundant than previously believed (Grindal et al. submitted). However, many confounding variables prevent these data from being useful in understanding relative abundance of northern myotis in
northeastern Alberta. Mistnetting efforts vary highly among bat biologists doing work in this region, with those using high nets strung across relatively large seismic cutlines producing a lower capture ratio of northern myotis than those using lower nets strung in more cluttered intact forest where few, if any, other northern Alberta bat species tend to forage (C. Lausen, pers. obs., D. Player, pers. comm.). This bias in mistnetting effort can have dramatic effects on relative species abundances in an area. For example, two mistnetting surveys that took place in southwest B.C. in exactly the same netting locations (sites were within metres of each other), each for two years (one survey: 2000 and 2001; subsequent survey: 2007 and 2008), resulted in 76 bat captures in 2000 and 2001, 10.5% of which were species of long-eared bats (either Keen’s myotis or long-eared myotis); yet in 2007 and 2008, 42.9% of 63 captures were long-eared species (T. Luszcz, unpubl. data, C. Lausen and D. Nagorsen, unpubl. data). The difference between these two survey efforts was that in 2007 and 2008, long-eared bat species were targeted and thus nets were placed lower and among more clutter than in 2000 and 2001.

Other confounding factors in understanding northern myotis abundance and distribution in northern Alberta is that most northern myotis and little brown myotis captures have not been genetically confirmed, leaving room for error in species identification. It is also common practice to mistnet survey areas within proposed development sites that are likely to provide good foraging/roosting opportunities for the only bat of conservation concern in northern Alberta, the northern myotis (May Be At Risk; Alberta Sustainable Resource Development 2007); in other words, targeting older deciduous stands in close proximity to water (C. Lausen, pers. obs.). As such, it is impossible to evaluate to what extent, if any, northern myotis occur outside of these targeted optimal habitat patches. It is also possible that targeting northern myotis (low nets in cluttered forest or narrow seismic/cutlines) could produce a capture ratio for bat species representative of habitat that is optimal for northern myotis, thus under-representing the other bat species in northern Alberta, which do not typically glean prey and as such are more open- and/or high-flying species (little brown myotis, long-legged myotis, silver-haired bat [Lasionycteris noctivagans], hoary bat [Lasiurus cinereus], eastern red bat [Lasiurus borealis]). Details of capture conditions and locations would be needed to properly determine any bias toward capture of northern myotis in environmental assessments of northern Alberta.

As a result of inconsistent use of genetic testing, patchy sampling of bats, variation in mistnetting techniques, and small sample sizes in most areas, it is not yet clear what the ratio of captures of this species is in relation to other northern Alberta bat species, but Table 1 provides a summary of projects in Alberta reporting capture of at least one northern myotis (captures from most areas and projects not confirmed by genetics). This is an overestimate of relative abundance given that it does not include capture efforts when no northern myotis were captured (i.e., if no northern myotis were captured, data were not reported). Capture ratios from some surveys outside Alberta where northern myotis were captured have been included to allow a crude comparison.

3. Population Trends. - We have no baseline data for bat population sizes in this province and thus no understanding of population trend as anthropogenic disturbance increases in northern Alberta. Sporadic sampling efforts in various locations of northern Alberta have taken place over the past 20 or more years (Table 1); some surveys have been of general bat biodiversity, others targeted certain species for research or environmental assessment. Despite potential inherent biases (see above), these are our only data upon which to infer relative abundance and distribution of northern myotis across central and northern Alberta.
Table 1. Summary by region of published studies reporting capture of the northern myotis (MYSE) in Canada. Identification of captures from most areas and projects were based on morphometrics only, and were not confirmed by genetic testing.

<table>
<thead>
<tr>
<th>General Location</th>
<th>Specific Locations Sampled</th>
<th>MYSE captures</th>
<th>Total bats captured</th>
<th>Total species captured</th>
<th>Sources</th>
</tr>
</thead>
<tbody>
<tr>
<td>North-central Alberta</td>
<td>Caribou River, Wabasca River, Lesser Slave Lake</td>
<td>2 (4%)</td>
<td>49</td>
<td>4</td>
<td>Vonhof and Hobson (2001)</td>
</tr>
<tr>
<td>Northwest Alberta</td>
<td>Rainbow Lake, Sousa Creek, Grande Prairie, EMEND Research Area northwest of Peace River</td>
<td>29 (33%)</td>
<td>89</td>
<td>3</td>
<td>Lippert (2001); Marinelli (2000); Patriquin (2001); Vonhof and Hobson (2001)</td>
</tr>
<tr>
<td>Northeast Alberta</td>
<td>Athabasca; Crow Lake; Anzac; Conklin; Fort MacKay; La Butte Wildland Provincial Park; Fidler-Greywillow Wildland Park; Lac La Biche; various other locations in the oil sands area near Fort McMurray</td>
<td>264 (38%)</td>
<td>697</td>
<td>5</td>
<td>Crampton (1995); Grindal et al. (submitted); Hubbs and Schowalter (2003)</td>
</tr>
<tr>
<td>Central Alberta</td>
<td>Whitecourt, Elk Island National Park, Cooking Lake - Blackfoot Recreation Area, Edmonton area, Drayton Valley, Content Bridge (on Red Deer River), Battle River (near Donalda and in Big Knife Prov. Park)</td>
<td>11 (5.5%)</td>
<td>199</td>
<td>6</td>
<td>Holroyd (1983); Lausen (2006a); Lippert (2001); Marinelli (2000); Vonhof and Hobson (2001)</td>
</tr>
</tbody>
</table>

**OUTSIDE OF ALBERTA:**

<table>
<thead>
<tr>
<th>SW Northwest Territories</th>
<th>Nahanni National Park</th>
<th>3 (17%)</th>
<th>18</th>
<th>4</th>
<th>Lausen (2006b)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Northeast B.C.</td>
<td>Fort Nelson and Highway 77 region</td>
<td>9 (23%)</td>
<td>39</td>
<td>4</td>
<td>Vonhof et al. (1997)</td>
</tr>
<tr>
<td>Southeastern B.C.</td>
<td>Revelstoke area</td>
<td>17 (15.5%)</td>
<td>110</td>
<td>7</td>
<td>Haney et al. (2002); Caceres (1998)</td>
</tr>
<tr>
<td>New Brunswick</td>
<td>in and around Fundy National Park</td>
<td>135 (49%)</td>
<td>277</td>
<td>2</td>
<td>Broders et al. (2006)</td>
</tr>
</tbody>
</table>

1 some or all *Myotis* samples were genetically tested for confirmation of species identification; all northern myotis were genetically confirmed.
Within some areas of its Alberta range, the northern myotis is captured in low ratios (Table 1). For example, 1 of 99 bats captured in the Lac La Biche area (Crampton 1995), 1 of 18 bats captured at Elk Island National Park (Holroyd 1983), 8 of 43 bats captured in the greater Edmonton area (including Elk Island National Park; Vonhof and Hobson 2001), 1 of 83 bats captured in the Red Deer and Battle Rivers Parkland zone (Lausen 2006a), and 4 of 443 bats captured in various central Alberta locations (M. Pybus, unpubl. data; not shown in Table 1) were identified as northern myotis. Substantial capture numbers of northern myotis (264 northern myotis out of a total of 697 captures) have come only from northeastern Alberta as part of numerous environmental impact assessments in the oil sands area around Fort McMurray. If species identifications are reliable (Grindal et al. [submitted] report 14% - 17% error in species identifications made by experienced field biologists), then northern myotis appear to be locally abundant in the oil sands area of northeastern Alberta (Grindal et al. submitted), at least in the sites targeted for sampling (before development) in this area. Interestingly, 2008 mistnetting surveys in various locations around Ft. McMurray produced a northern myotis capture ratio of 6.2% (5 of 81 bat captures were northern myotis; D. Player, unpubl. data). This is in stark contrast to the 38% northern myotis ratio reported in this area from 1995 - 2006 (Table 1), and may reflect local differences in habitat sampled, weather, or differences in mistnetting techniques, given that most bat captures in the 2008 surveys were in a relatively new commercially available tall netting system using 7.4 m poles (“triple high net”; Bat Conservation and Management, Pennsylvania) set in fairly wide open cutlines (C. Lausen, pers. obs., D. Player, unpubl. data).

Nearly 30 years ago, northern myotis made up a small proportion of the individuals found hibernating in both Cadomin Cave (Schowalter 1980) and in Wood Buffalo National Park (Schowalter 1979). Recent surveys of these hibernacula have been point counts without species identifications; therefore, nothing is known of recent relative abundance at these hibernacula.

We cannot definitively say what impact continued industrial development in northeastern Alberta may have on northern myotis populations; however, given their dependence on forest interiors, any development that reduces the present and future density of older stage (“over-mature”) trees is likely to reduce the population size of this species.

**LIMITING FACTORS**

1. **Availability of Summer Roosting Habitat.** - There are an increasing number of studies on roosting that find northern myotis roosting in fragmented forested environments. Although these studies provide evidence that northern myotis roost in fragmented forests, we do not have any long-term data to indicate if populations are or are not viable over the long term in these disturbed ecosystems. Recent studies in fragmented forest environments found that northern myotis decrease movement and thereby shrink home ranges size in response to removal of intact forest habitat (Foster and Kurta 1999, Henderson and Broders 2008), which likely reflects limitation of roost and foraging resources, and may ultimately result in decreased population density. In highly fragmented areas of Prince Edward Island, mean movement distance between capture and first-located roosts of northern myotis was 285 m (Henderson and Broders 2008), which is much smaller than the 1001 m average movement distance reported in intact forests of New Brunswick (Broders et al. 2006). Foraging areas in intact forests have been reported to be between 46 ha and 65 ha (Broders et al. 2006, Owen et al. 2003), in stark contrast to 6 ha in fragmented areas (Henderson and Broders 2008). Because these bats do not fly in open areas (Henderson and...
Broders 2008, Patriquin and Barclay 2003), it is not surprising to find that home ranges shrink as intact forest fragment size decreases. This corresponding lower availability of resources for roosting and foraging is likely to result in patches accommodating fewer bats (because of emigration or lower reproductive rates; Ricklefs 1990).

Unfortunately, there have been no population size estimates, and no studies of survival and fecundity of northern myotis to demonstrate the impact of industrial activities on this species in highly fragmented ecosystems. As a forest-dependent species, northern myotis will have an optimal intact forest patch size, and anything smaller will inevitably compromise roosting and foraging opportunities, thereby decreasing individual fecundity and ultimately population size. It is likely that this minimum intact forest patch size varies across the species’ range with local differences in insect, water and roost resources. On Prince Edward Island, Henderson et al. (2008) found the smallest forest fragments used by northern myotis were 17.1 ha and 15.9 ha for females and males, respectively. Studies of northern myotis in areas where intact forest still exists adjacent to fragmented forest find northern myotis preferentially use intact forest (Carter and Feldhamer 2005, Owen et al. 2003).

There are mixed results in the literature regarding type of tree roosts selected by northern myotis. Some roosting studies of this species conclude that they depend on mature/old growth forests (as summarized in Caceres and Pybus 1997), with bats selecting large diameter tall roost trees (Sasse and Pekins 1996, Vonhof and Wilkinson 1999); other studies conclude that the northern myotis is more of a tree-roosting generalist, roosting in trees of various height, diameter, and decay stages (e.g., Carter and Feldhamer 2005, Foster and Kurta 1999, Vonhof et al. 1997). Additionally, it is not clear whether stand age is important for roost plots; in a study in South Dakota, a few northern myotis were documented roosting in remnant snags in recently forested plots (< 5 years; Cryan et al. 2001), although the largest maternity colony of northern myotis in this study ($n = 41$) was found in a nearby mature forest stand that had not been harvested in more than 50 years (Cryan et al. 2001). Many studies implicate the importance of snag density and stem density in roost plots, suggesting relatively abundant snags and dense forest are habitat requirements for this species (Cryan et al. 2001, Foster and Kurta 1999, Menzel et al. 2002, Owen et al. 2002, 2003).

A common thread among roost studies for northern myotis seems to be the preference for deciduous trees, especially for females (e.g., Broders and Forbes 2004, Henderson and Broders 2008, Henderson et al. 2008). Studies of northern myotis in agriculture-forest mosaic landscapes (eastern Canada and Michigan) found that females of this species roost exclusively in deciduous trees (Foster and Kurta 1999, Henderson and Broders 2008). However, the importance of coniferous trees in some locations has also been reported (e.g., Lacki and Schwierjohann 2001), especially for males (Broders and Forbes 2004, Henderson et al. 2008). This suggests that mixedwood stands are important for northern myotis because of intersexual variation in habitat selection. In the past, forest policy in Alberta generally favoured exploitation and subsequent elimination of mixedwood stands, with reforestation to pure deciduous or pure conifer stands typical (Province of Alberta 1991). Modeling of northern Alberta forest structure based on a variety of these harvesting practices predicted between 0% and 10% old-growth deciduous and between 0% and 23% old-growth coniferous remaining in 300 years, with the conclusion that projected levels of tree harvesting would be sustainable, but many wildlife species would lose large amounts of preferred habitat (Cumming et al. 1994). Some current Forest Management Plans suggest a move toward trying to maintain more mixedwood stands and
retention of a greater percentage of old-growth trees (e.g., Alberta-Pacific 2006).

On Prince Edward Island, Henderson and Broders (2008) suggested that radio-tracked northern myotis foraged along forested riparian areas rather than in the surrounding cleared areas, not because of prey availability, but because of a strong preference for foraging and commuting within forested landscapes. The presence of this species in highly fragmented landscapes suggests northern myotis could be more resilient to some degree of forest fragmentation than once believed; however, there is a need to conduct roosting and foraging studies of this species in Alberta.

Presumably, this species has evolved with a natural disturbance regime, with fire creating a mosaic of live and dead trees in clumped patches (Hansson 1992, EMEND 2006). Bats may be able to shift their behaviour in response to small habitat disturbances (Patriquin and Barclay 2003); however, in large scale disturbances such as large clearcuts or fires with salvage logging, few over-mature (older than typically required for harvesting), dying or dead trees remain (Hansson 1992) to be used as bat roosts. Furthermore, changes in the density of vegetation will influence insect abundance and foraging opportunities, especially for gleaning bat species, such as northern myotis, who prefer and seemingly depend on vegetatively cluttered habitat. Patriquin and Barclay (2003) found in Alberta that clear-cutting of forests resulted in increased activity of larger species of bats that typically forage away from clutter (i.e., silver-haired bat), but decreased activity of northern myotis. This likely reflects their gleaning behaviour and tendency to stay within intact forest. Forestry activities should strive to mimic a natural disturbance regime, leaving a mosaic of interconnected forest patches varying in tree density (Walker et al. 1995, Patriquin and Barclay 2003). Unfortunately, the large-scale hot-burning fires that more typically occur today where fires have been historically suppressed tend to clear large tracts of forest rather than leave a patch mosaic of living and dead trees (Agee 1996). An example of a severe forest fire in northern Alberta is the House Fire of 2002, which burned 250,000 ha of forest in northeastern Alberta; salvage logging occurred over much of this area (Macdonald 2006).

The Alberta boreal forest covers a vast area, but how much of this forest is usable roosting habitat for northern myotis is not known. Just as the forest harvesting industry recognizes that not all of the boreal forest contains trees usable for timber (e.g., only one third of Al-Pac Forest Products, Inc. Forest Management Area contains trees usable for timber; Alberta-Pacific 2006), much of the boreal area may be unsuitable for northern myotis. Once roosting requirements for this species in Alberta are understood, we will know what percentage of the boreal forest is currently suitable habitat for this species.

2. Hibernacula - The availability of suitable hibernacula may be a limiting factor in northern myotis populations. Natural caves or abandoned mines may be in short supply over much of the species’ range in northern Alberta. In addition, environmental limitations that govern successful hibernation make many mines and caves unsuitable (Ministere de l’Environnement et de la Faune 1996).

Hibernating bats are particularly sensitive to disturbance (Thomas 1995). They have limited stored energy supplies and have no opportunity to replace energy expended during the winter. To conserve energy, northern myotis, like many bat species, are true hibernators. That is, they enter a state of torpor in which their internal body temperature approaches freezing, and their breathing and heart rate are significantly slowed. Bats will arouse from torpor to seek water, or move to another location if environmental conditions become unsuitable (for example, if the hibernaculum becomes too warm or cold), or if they are physically disturbed. Most often this
latter disturbance is related to human activity within the hibernaculum (Thomas 1995).

In general, bat hibernation is poorly understood. Bats arouse periodically throughout the winter for various reasons, and may fly, although why flight occurs is not yet understood given its energetic expense to a hibernating bat (Lausen and Barclay 2006a). Each time a bat arouses from torpor, it uses significant energy supplies to warm its body and increase its metabolic rate. Arousals generally account for 80% to 90% of the energy expenditure in hibernating animals during the winter (Thomas et al. 1990). Thus, if bats repeatedly arouse during the winter, they may expend all their energy supplies and die. Human disturbances in hibernacula have the potential to cause some bats to arouse more frequently than can be supported by their fat reserves.

Hibernating bats are more sensitive to human presence than was previously supposed. Thomas (1995) documented arousal and flight of hibernating bats within 30 minutes of humans entering a hibernaculum of little brown myotis and northern myotis. The disturbing stimulus appeared to be the mere presence of humans and their associated lights and sounds. At no time did any of the humans touch or closely approach the hibernating bats. Further, increased bat activity continued for up to seven or eight hours after the humans had left the cave. This implies that aroused bats may have been disturbing torpid bats, resulting in other bats arousing. Thus, the original disturbance by humans is magnified among the hibernating bats. In some cases it may be desirable to prevent unnecessary disturbance by blocking the entrance to a hibernaculum. Barred gates may be used to allow entry only by bats and authorized humans or to limit access to times when bats are not at risk (for example, summer). However, there are some inherent risks with this approach, such as the potential destruction of the gate and surrounding cave entrance by humans wishing to regain access to the cave/mine (C. Lausen, pers. obs.), and the potential alteration of airflow, cave microclimate, and the cave ecosystem; furthermore, gates may make access to the cave by various wildlife, including bats, more difficult (Elliot 2006, Martin et al. 2000). Gating should be properly assessed to determine its necessity, appropriate structure, and placement; all aspects of access management need to be carefully evaluated before using gates as a tool (Elliot 2006).

The abundance of the northern myotis may be limited by closing or otherwise altering old mines or caves that are essential for hibernation. Changes to the internal environment of a hibernaculum may also result in its abandonment by bats. Although there are no records of closures of known bat hibernacula in Alberta, such situations have occurred elsewhere (Taylor 1995, 1996, Tuttle 1996). Furthermore, our lack of knowledge regarding the hibernation sites of the northern myotis in Alberta could lead to unintentional disturbance or destruction of hibernacula. Because it has recently been shown that some species of bats in Alberta hibernate in deep rock crevices along river valley walls (Lausen and Barclay 2006a), and not necessarily in caves, it is possible that a large number of bat hibernation sites have been overlooked and could easily be destroyed or disturbed.

In addition to providing critical winter habitat, hibernacula are also important as sites of swarming (breeding behaviour) in many bat species. Many bats that swarm at a site do not hibernate there (Fenton 1983). This is a critical period and the only time when large numbers of male and female bats occur in the same place at the same time. In Alberta, swarming of Myotis spp. at Cadomin Cave occurs in late August and early September (Schowalter 1980). Although a few northern myotis were caught there, it is not known if Cadomin Cave is a primary swarming site for this species. Protection of swarming sites is particularly important for breeding purposes, and has
implications beyond the group of bats that use the specific hibernaculum. Alteration of the site or disturbance of the bats during swarming should be kept to a minimum.

3. White Nose Syndrome. - When the number of hibernacula is low, bat populations are disproportionately clumped at a time when they are most vulnerable. This increases the risk that a catastrophic event may seriously reduce the population. Such a catastrophic event has taken place recently in cave hibernacula in the northeastern United States. White Nose Syndrome (WNS) is a recent phenomenon describing the occurrence of a white fungus on hundreds of thousands of dead bats during hibernation in 2006-2007 and 2007-2008. This fungus thrives at cold temperatures, propagating in bat tissue when bats are in torpor with suppressed immune function (Blehert et al. 2008). Fungal hyphae fill hair follicles, sebaceous glands and sweat glands, eroding the epidermis of ears and wings, which likely stimulates an arousal by bats to boost immune function; most bats found dead with this fungus had little or no identifiable fat reserves, suggesting starvation (Blehert et al. 2008).

Currently, WNS is found only in the northeastern United States, and its long distance spread from one to five states in one winter season (2007 – 2008) led the US Fish and Wildlife Service to issue containment and decontamination protocols for bat biologists and cavers visiting caves in the affected and neighbouring states (U.S. Fish and Wildlife Service 2008). Winter survey counts of all bat species in affected cave hibernacula show declines ranging from 81% to 97% (Hicks et al. 2008). WNS in bats has been likened to chytridiomycosis, a lethal fungal skin infection associated with recent worldwide declines of amphibians (as summarized in Blehert et al. 2008). It is suspected that bats transfer the fungal spores of this syndrome to neighbouring locations; however, it is likely that humans (e.g., cavers, bat biologists) can propagate the spread of this fungus unless precautions are taken (U.S. Fish and Wildlife Service 2008). It is forecast that WNS may spread into southern Canada during the winter of 2008 – 2009 (CBC 2008). Although little is known about bat hibernacula in Alberta, it is possible that appropriate microclimatic conditions could exist for this fungus to affect Alberta bats, should the spores reach this province. It is speculated that spores can lie dormant, but remain viable, in inappropriate (i.e., warm, dry) conditions, allowing the spread of the fungus through inhospitable habitats or seasons (D. Blehert, pers. comm.).

It has recently been predicted that if WNS continues at its current rate of spread and fatality, that little brown myotis, northern myotis and tri-coloured bat (Eastern pipistrelle; Perimyotis subflavus) will be extirpated from eastern North America (Hicks et al. 2008). Only time will tell whether WNS represents a threat to cave-hibernating species in western North America, and whether the western populations of northern myotis will have elevated importance on a continental scale as eastern populations continue to decline. Because so few Myotis hibernacula are known in Alberta, and bat population numbers are unknown, it is possible that if WNS were to affect bats in this province, its presence would go unnoticed for quite some time, evident only when highly scarred wing tissues are noticed or if drastic declines in summer populations of cave-hibernating species are documented.

4. Climate Change. - In general, bat distributions and abundance in Canada tend to be limited by climate and availability of suitable roost sites (van Zyll de Jong 1985). Overlying this trend is a pattern of decreased diversity and abundance of bats with increasing latitude north of the equator. As such, in the northern temperate regions of Canada, the number of bat species and individuals is relatively low. The most northerly distribution tends to coincide with the limit of mean annual growing season of 120 days; that is, the period during which
the mean daily temperature remains above 6°C (van Zyll de Jong 1985). This appears to reflect the minimum time needed for female bats to produce young and have them successfully grow and develop to the stage where they can survive hibernation. Thus, climate may be the ultimate limiting factor on the northern distribution of northern myotis. Their distribution may shift north with climate change (Humphries et al. 2002), although other factors, such as availability of roosts and length of nights are likely to limit northern expansion.

Recently it has been determined that climate change has increased the mortality rate of trees across the western United States, with mortality rates doubling over a 17-year period in Pacific Northwest forests, and over a 29-year period in Interior forests (van Mantgem et al. 2009). Slightly warmer and drier conditions have been implicated as the cause of this increased mortality, reduced recruitment, and changed forest structure and composition discovered in long term datasets of unmanaged old growth forests. All sizes and species of trees are experiencing increased mortality across the west. Van Mantgem et al. (2009) point out that a persistent doubling of tree mortality rate in old-growth forests would ultimately cause a greater than 50% reduction in average tree age in these undisturbed forests and hence a potential overall reduction in mean tree size. As these findings are new, it is not yet clear what impact this will have on northern myotis and other tree-dwelling bat species in the long term, but if density and size of trees in old-growth stands continue to decrease, roosting and foraging habitats will inevitably be altered.

**STATUS DESIGNATIONS**

1. **Alberta.** - In 1991, the general status of the northern myotis was considered Undetermined because of the lack of information on the biology and population trends of this species (Alberta Fish and Wildlife Division 1991). However, in 1996, the species was moved to the “Blue List” of species that may be at risk in the province (Alberta Environmental Protection 1996). This change in status was based on the relative rarity of the species in the province, and on its apparent reliance on mature trees for roosting. Since 2000, this general status category has been renamed as May Be At Risk in Alberta (Alberta Sustainable Resource Development 2001, 2007).

Although the provincial general status of May Be At Risk offers no legal protection for those wildlife species, hibernacula of the northern myotis, and all other bats in Alberta, are afforded protection from disturbance between September 1 and April 30 under Section 38 (1) of Alberta’s Wildlife Act.

As a group, bats are identified in the Wildlife Act (1986) as “non-licence” species. This designation means that bats can be hunted or harvested without a permit. However, it is not legal to possess live bats in Alberta (Section 3, Captive Wildlife Regulation of Alberta’s Wildlife Act) because of the concern for public safety. A permit is required to handle bats for research purposes (Wildlife Regulation 143/97, ss. 81, 82).

The Alberta Natural Heritage Information Centre ranks the northern myotis as S2S3 (tracking hibernacula only) because of concern for roosting and breeding habitat (Alberta Natural Heritage Information Centre 2007, 2008).

2. **Other Areas.** - Given its broad range and relatively abundant populations in the east, northern myotis has a Canadian general status rank of Secure, which changed from its rank of Sensitive in 2000 (CESCC 2006). Quebec is currently the only state/province in the North American distribution of this species that lists

* See Appendix 1 for definitions of selected status designations.
northern myotis as Secure (NatureServe 2008). If WNS spreads to eastern Canada from eastern U.S., as predicted, this Secure designation will most likely change.

The northern myotis is on the “Blue List” (special concern – at risk but not threatened) in British Columbia (B.C. Conservation Data Centre 2008). This species was recently discovered in the Yukon (Jung et al. 2006; Lausen et al. 2008), where it was given a May Be At Risk general status rank (T. Jung, pers. comm.). Northern myotis was known from only a few specimens in the Northwest Territories prior to 2006 when this species made up 17% of the captures in a formal survey of Nahanni National Park Reserve in the southwestern part of the territory (Lausen 2006b). It currently has not been ranked in the Northwest Territories (Status Undetermined; Working Group on General Status of NWT Species 2006). In the Maritimes, where the most research has occurred on this species in Canada, northern myotis is ranked as Sensitive (CESCC 2006). In forested areas, this species is captured nearly as often as the little brown myotis in New Brunswick (see Table 1; Broders et al. 2006), and in greater numbers than the little brown myotis in Nova Scotia (Broders et al. 2003); its ranking as Sensitive in these areas stems from its dependency on forests, a heavily exploited resource in many areas of Canada.

NatureServe classifies this species as S2S3 (imperilled/vulnerable) in B.C., Yukon and Montana (NatureServe 2008). In Saskatchewan, the breeding population is considered S4 (apparently secure), while the nonbreeding populations remain unranked (SNRN; NatureServe 2008). In Manitoba, the breeding population is also considered S4 and the nonbreeding population is considered S3S4. Globally, NatureServe ranks the northern myotis as G4 (apparently secure, last reviewed 19 April 2005; NatureServe 2008).

**RECENT MANAGEMENT IN ALBERTA**

Other than baseline efforts to document detailed status (Caceres and Pybus 1997), no management or research activities specifically focused on northern myotis have taken place in Alberta.

**SYNTHESIS**

An accurate assessment of the status of a species requires an understanding of its basic biology, ecology, and population size and trends over time. For the northern myotis, we have some basic information concerning its life history and habitat preferences, and can make limited inferences from knowledge of closely related species to fill in some information gaps. However, there are no good baseline data upon which to estimate population size or deduce/predict trends. Confusion over species identification has been a complicating factor in understanding the distribution and relative abundance of this species. While this species was likely underestimated in early bat studies of northern Alberta, it is now equally possible that it is overestimated if ear length and pelage colour are the only criteria used to differentiate this species from other similar Myotis. Acoustically, northern myotis is difficult to differentiate from other Myotis because of overlap of call characteristics, and any attempt to do so is likely to either underestimate or overestimate its abundance, depending on criteria used. DNA sampling of captures with subsequent genetic sequencing is the only way to ensure correct species identification. Special effort needs to be made to set mistnets that are most suitable for catching this species (i.e., nets that are set in vegetatively cluttered areas and extend to the ground and to vegetative edges). To best estimate species ratios, variation in nets needs to be used to maximize the likelihood of catching all possible species (i.e., tall nets in addition to nets that extend to the ground; nets in vegetatively cluttered areas and in open...
areas; nets over water; nets on narrow and wide trails/cutlines and in intact forest).

Capture records suggest northern myotis occur at varying densities across a broad area of northern Alberta. This species requires intact forest stands for roosting and foraging during the summer; however, patch sizes, and preferred roost tree and stand characteristics are not known for Alberta. Caves appear to be the preferred hibernacula, but other sources of hibernacula such as rocky riparian areas have not been investigated. Only two caves in Alberta are known to contain hibernating northern myotis.

Recent roosting and foraging studies from outside of Alberta and Canada have shed light on the ecology of this species, but have also produced some conflicting findings. Even though northern myotis are increasingly being found in young forests, population viability studies have yet to be conducted. Finding northern myotis in these habitats, which were not previously associated with this species but are increasingly available, may be indicative of plasticity in behaviour or merely philopatry (returning to place of birth). Long-term persistence of this species in disturbed ecosystems has yet to be documented. Local studies, specific to Alberta, are likely to be more informative than extrapolation of results from other ecosystems. More extensive and intensive mistnetting surveys are needed in Alberta; only the oil sands area near Fort McMurray has been extensively surveyed, with relatively large numbers of northern myotis being reported from this northeastern Alberta location. Most of these northeastern Alberta surveys have been done as part of environmental assessments prior to industrial developments. It is not clear whether post-development surveys would yield similar capture rates of the species; similarly, because of a lack of bat research in Alberta’s Boreal Forest Natural Region, habitat requirements of northern myotis and response to forest fragmentation are unknown, making it unclear what the long-term population trend might be for northern myotis in these increasingly developed industrial regions.

The Alberta boreal forest covers a vast area, and the natural patchiness of suitable intact forest habitat suggests northern myotis are likely to be patchily distributed. It is difficult to accurately assess the proportion of Alberta’s boreal forest that can be considered suitable habitat for this species, or the impacts that may result from various resource extraction activities (such as forestry), knowing so little about the habitat requirements for this species in Alberta. Because summer roosting requirements have been shown to differ between males and females, and the distribution of males and females seems to be affected by fragmentation at different scales (forest fragment scale for females, landscape scale for males), roosting studies of both sexes are required so that strategies for viable northern myotis populations are written into forest management plans.

Roost selection criteria for both males and females of this species in Alberta (at three scales: tree, stand, and landscape) should be considered a priority research item. A habitat-based model for Alberta should follow, in order to determine the degree of threat of industrial activities to the northern myotis bat in Alberta. Habitat suitability models in New Brunswick have shown a preference for deciduous-dominated stands by females, and a preference for coniferous-dominated stands by males. In Alberta, because we do not know roost preferences, we cannot predict what impact industrial activities will have on male and female northern myotis, the degree of fragmentation northern myotis may tolerate in the boreal ecosystem, and what minimum and optimal patch sizes of intact forest (deciduous, coniferous, mixedwood) are required to maintain a viable population. In addition, more hibernation and breeding sites need to be located, not only to avoid accidental destruction of such critical habitat resources,
but also to ensure that forest patch dynamics facilitate necessary movement patterns by individuals. Given that this species is thought to remain only in forest interiors, commuting within the forest or along forest edges, but not through open environments (Hogberg et al. 2002), it would be critical to accommodate nightly movements for drinking, foraging, and roosting, and seasonal movements for mating and hibernating. Only with these data could appropriate forest management plans be designed to ensure the persistence of healthy populations of this forest-dependent species in Alberta.

Minimal survey effort has gone into surveying bats in northwestern Alberta; more inventory work should be done here, as well as in north-central and central Alberta, to understand whether the northern myotis is common across its range in Alberta, or patchily distributed. All bat surveys should include intensive research programs, including radiotelemetry studies, which would allow identification of male and female roost sites and a quantitative assessment of the year-round habitat requirements and preferences of this species. Such information would not only enhance our ability to assign an appropriate status ranking for this species, but it would provide data for design of appropriate forest management plans, and clarify potential factors that might limit the abundance and distribution of the northern myotis in the province.
LITERATURE CITED


Barclay, R.M.R., J. Ulmer, C.J.A. MacKenzie, M.S. Thompson, L. Olson, J. McCool,


Lippert, H.D. 2001. The relationship of bat activity to habitat type and structure retention in managed, Boreal forests of West-Central Alberta. MSc. thesis, Department of Renewable Resources, University of Alberta, Edmonton, AB.


Appendix 1: Definitions of status ranks and legal designations.


<table>
<thead>
<tr>
<th>2005 Rank</th>
<th>1996 Rank</th>
<th>Definitions</th>
</tr>
</thead>
<tbody>
<tr>
<td>At Risk</td>
<td>Red</td>
<td>Any species known to be <em>At Risk</em> after formal detailed status assessment and designation as <em>Endangered</em> or <em>Threatened</em> in Alberta.</td>
</tr>
<tr>
<td>May Be At Risk</td>
<td>Blue</td>
<td>Any species that may be at risk of extinction or extirpation, and is therefore a candidate for detailed risk assessment.</td>
</tr>
<tr>
<td>Sensitive</td>
<td>Yellow</td>
<td>Any species that is not at risk of extinction or extirpation but may require special attention or protection to prevent it from becoming at risk.</td>
</tr>
<tr>
<td>Secure</td>
<td>Green</td>
<td>Any species that is not <em>At Risk</em>, <em>May Be At Risk</em> or <em>Sensitive</em>.</td>
</tr>
<tr>
<td>Undetermined Status</td>
<td>Undetermined</td>
<td>Any species for which insufficient information, knowledge or data is available to reliably evaluate its general status.</td>
</tr>
<tr>
<td>Not Assessed</td>
<td>n/a</td>
<td>Any species that has not been examined during this exercise.</td>
</tr>
<tr>
<td>Exotic/Alien</td>
<td>n/a</td>
<td>Any species that has been introduced as a result of human activities.</td>
</tr>
<tr>
<td>Extirpated/Extinct</td>
<td>n/a</td>
<td>Any species no longer thought to be present in Alberta (Extirpated) or no longer believed to be present anywhere in the world (Extinct).</td>
</tr>
<tr>
<td>Accidental/Vagrant</td>
<td>n/a</td>
<td>Any species occurring infrequently and unpredictably in Alberta, i.e., outside its usual range.</td>
</tr>
</tbody>
</table>

B. Alberta Species at Risk Formal Status Designations
Species designated as *Endangered* under Alberta’s *Wildlife Act* include those listed as *Endangered* or *Threatened* in the Wildlife Regulation (in bold).

<table>
<thead>
<tr>
<th>Status</th>
<th>Definition</th>
</tr>
</thead>
<tbody>
<tr>
<td>Endangered</td>
<td>A species facing imminent extirpation or extinction.</td>
</tr>
<tr>
<td>Threatened</td>
<td>A species likely to become endangered if limiting factors are not reversed.</td>
</tr>
<tr>
<td>Species of Special Concern</td>
<td>A species of special concern because of characteristics that make it particularly sensitive to human activities or natural events.</td>
</tr>
<tr>
<td>Data Deficient</td>
<td>A species for which there is insufficient scientific information to support status designation.</td>
</tr>
</tbody>
</table>

C. Committee on the Status of Endangered Wildlife in Canada (after COSEWIC 2006)

<table>
<thead>
<tr>
<th>Status</th>
<th>Definition</th>
</tr>
</thead>
<tbody>
<tr>
<td>Extinct</td>
<td>A species that no longer exists.</td>
</tr>
<tr>
<td>Extirpated</td>
<td>A species that no longer exists in the wild in Canada, but occurs elsewhere.</td>
</tr>
<tr>
<td>Endangered</td>
<td>A species facing imminent extirpation or extinction.</td>
</tr>
<tr>
<td>Threatened</td>
<td>A species that is likely to become endangered if limiting factors are not reversed.</td>
</tr>
<tr>
<td>Special Concern</td>
<td>A species that may become threatened or endangered because of a combination of biological characteristics and identified threats.</td>
</tr>
<tr>
<td>Not at Risk</td>
<td>A species that has been evaluated and found to be not at risk of extinction given the current circumstances.</td>
</tr>
<tr>
<td>Data Deficient</td>
<td>A category that applies when the available information is insufficient to (a) resolve a wildlife species' eligibility for assessment, or (b) permit an assessment of the wildlife species' risk of extinction.</td>
</tr>
</tbody>
</table>
### D. Heritage Status Ranks: Global (G), National (N), Sub-national (S) (after Alberta Natural Heritage Information Centre 2007, NatureServe 2008)

<table>
<thead>
<tr>
<th>Rank</th>
<th>Description</th>
</tr>
</thead>
<tbody>
<tr>
<td>G1/N1/S1</td>
<td>5 or fewer occurrences or only a few remaining individuals. May be especially vulnerable to extirpation because of some factor of its biology.</td>
</tr>
<tr>
<td>G2/N2/S2</td>
<td>6 to 20 or fewer occurrences or with many individuals in fewer locations. May be especially vulnerable to extirpation because of some factor of its biology.</td>
</tr>
<tr>
<td>G3/N3/S3</td>
<td>21 to 100 occurrences; may be rare and local throughout its range, or in a restricted range (may be abundant in some locations). May be susceptible to extirpation because of large-scale disturbances.</td>
</tr>
<tr>
<td>G4/N4/S4</td>
<td>Typically &gt; 100 occurrences. Apparently secure.</td>
</tr>
<tr>
<td>G5/N5/S5</td>
<td>Typically &gt; 100 occurrences. Demonstrably secure.</td>
</tr>
<tr>
<td>GX/NX/SX</td>
<td>Believed to be extinct or extirpated; historical records only.</td>
</tr>
<tr>
<td>GH/NH/SH</td>
<td>Historically known; may be relocated in the future.</td>
</tr>
<tr>
<td>G?/N?/S?</td>
<td>Not yet ranked, or rank tentatively assigned.</td>
</tr>
</tbody>
</table>

### E. United States Endangered Species Act (after National Research Council 1995)

<table>
<thead>
<tr>
<th>Category</th>
<th>Description</th>
</tr>
</thead>
<tbody>
<tr>
<td>Endangered</td>
<td>Any species that is in danger of extinction throughout all or a significant portion of its range.</td>
</tr>
<tr>
<td>Threatened</td>
<td>Any species that is likely to become an endangered species within the foreseeable future throughout all or a significant portion of its range.</td>
</tr>
</tbody>
</table>
Appendix 2. Site locations of the northern myotis in Alberta. Map numbers refer to points in Figure 1. Non-published data were obtained from Fish and Wildlife Management Information System, Alberta Sustainable Resource Development (May 15, 2008 for data ≥1997, and Jan. 26, 2009 for data <1997).

<table>
<thead>
<tr>
<th>MAP #</th>
<th>LOCATION</th>
<th>SOURCE/MUSEUM RECORD NUMBER</th>
<th>Latitude (°N)</th>
<th>Longitude (°W)</th>
</tr>
</thead>
<tbody>
<tr>
<td>2</td>
<td>Hinton</td>
<td>M. Pybus (pers. comm., 1 bat banded and released in 1990)</td>
<td>53.67</td>
<td>-118.00</td>
</tr>
<tr>
<td>3</td>
<td>Edson</td>
<td>M. Pybus (pers. comm., 3 bats banded and released in 1991); PMA -- 83.23.9</td>
<td>53.58</td>
<td>-116.43</td>
</tr>
<tr>
<td>4</td>
<td>Edson area</td>
<td>University of Alberta Museum of Zoology(^2) field No. 160 (date not available)</td>
<td>53.58</td>
<td>-116.42</td>
</tr>
<tr>
<td>5</td>
<td>Lesser Slave Lake</td>
<td>PMA / 83.23.8 (1981)</td>
<td>55.50</td>
<td>-115.68</td>
</tr>
<tr>
<td>6</td>
<td>Drayton Valley</td>
<td>PMA / Z81.20.1 (1980); Lippert (2001)</td>
<td>53.00</td>
<td>-115.00</td>
</tr>
<tr>
<td>7</td>
<td>Spruce Grove</td>
<td>PMA / Z76.101.1 (1976)</td>
<td>53.55</td>
<td>-113.90</td>
</tr>
<tr>
<td>8</td>
<td>Edmonton</td>
<td>PMA -- Z76.123.1/Z78.69.1/Z79.87.1/83.23.7 (1976 - 1981)</td>
<td>53.53</td>
<td>-113.53</td>
</tr>
<tr>
<td>9</td>
<td>Elk Island National Park</td>
<td>Holroyd (1983); Vonhof and Hobson (2001)</td>
<td>53.62</td>
<td>-112.87</td>
</tr>
<tr>
<td>10</td>
<td>Lac La Biche area</td>
<td>Crampton (1995)</td>
<td>54.50</td>
<td>-112.00</td>
</tr>
<tr>
<td>11</td>
<td>Fort MacKay area</td>
<td>PMA / 83.22.1 (1983)</td>
<td>57.17</td>
<td>-111.62</td>
</tr>
<tr>
<td>14</td>
<td>Fort MacKay area</td>
<td>Environmental Impact Assessment for the Heavy Oil Lease Area, North of Fort McMurray, 2006, C. Godwin-Sheppard, pers. comm.</td>
<td>57.31</td>
<td>-110.50</td>
</tr>
<tr>
<td>15</td>
<td>NE of Ft. MacKay</td>
<td>Environmental Impact Assessment for the Heavy Oil Lease Area, North of Fort McMurray, 2006, B. Hamilton, pers. comm.</td>
<td>57.56</td>
<td>-110.70</td>
</tr>
<tr>
<td>16</td>
<td>Peace River</td>
<td>PMA / Z77.144.1 (1977)</td>
<td>56.22</td>
<td>-117.27</td>
</tr>
</tbody>
</table>
Appendix 2 continued:

<table>
<thead>
<tr>
<th>MAP #</th>
<th>LOCATION</th>
<th>SOURCE/MUSEUM RECORD NUMBER</th>
<th>Latitude (°N)</th>
<th>Longitude (°W)</th>
</tr>
</thead>
<tbody>
<tr>
<td>17</td>
<td>~70 km NW of Peace River</td>
<td>Patriquin (2001)</td>
<td>56.67</td>
<td>-118.00</td>
</tr>
<tr>
<td>18</td>
<td>Grande Prairie</td>
<td>PMA / Z78.123.1 (1978); Lippert (2001)</td>
<td>55.00</td>
<td>-119.00</td>
</tr>
<tr>
<td>19</td>
<td>Spirit River</td>
<td>PMA / 83.23.6 (1981)</td>
<td>55.78</td>
<td>-118.83</td>
</tr>
<tr>
<td>20</td>
<td>High Level</td>
<td>PMA / Z75.106.1 (1975)</td>
<td>58.52</td>
<td>-117.08</td>
</tr>
<tr>
<td>21</td>
<td>Wood Buffalo National Park</td>
<td>National Museum of Canada’s 43405 (date not available)</td>
<td>59.78</td>
<td>-112.32</td>
</tr>
<tr>
<td>22</td>
<td>Wood Buffalo National Park</td>
<td>PMA -- Z78.49.2 (1978)</td>
<td>59.67</td>
<td>-112.32</td>
</tr>
<tr>
<td>23</td>
<td>Wabasca River</td>
<td>Vonhof and Hobson (2001)</td>
<td>57.79</td>
<td>-115.34</td>
</tr>
<tr>
<td>24</td>
<td>Rainbow Lake</td>
<td>Vonhof and Hobson (2001)</td>
<td>58.30</td>
<td>-119.29</td>
</tr>
<tr>
<td>25</td>
<td>Sousa Creek</td>
<td>Vonhof and Hobson (2001)</td>
<td>58.68</td>
<td>-118.68</td>
</tr>
<tr>
<td>26</td>
<td>Conklin area</td>
<td>Hubbs and Schowalter (2003)</td>
<td>55.66</td>
<td>-111.14</td>
</tr>
<tr>
<td>27</td>
<td>Conklin area</td>
<td>StatOil Baseline Survey, 2008, Matrix Solutions, D. Player, pers. comm.</td>
<td>55.81</td>
<td>-111.33</td>
</tr>
<tr>
<td>28</td>
<td>Fort McMurray area</td>
<td>AOSC Baseline Survey, 2008, Matrix Solutions, D. Player, pers. comm.</td>
<td>56.69</td>
<td>-111.98</td>
</tr>
<tr>
<td>29</td>
<td>Fort McMurray area</td>
<td>Nexen Baseline Survey, 2008, Matrix Solutions, D. Player, pers. comm.</td>
<td>56.41</td>
<td>-110.86</td>
</tr>
<tr>
<td>30</td>
<td>Athabasca Tar Sands area</td>
<td>Suncor Voyageur Baseline Survey, 2004, Golder Associates, N. McDonald, pers. comm.</td>
<td>56.95</td>
<td>-111.51</td>
</tr>
<tr>
<td>31</td>
<td>Cooking Lake Blackfoot Provincial Park</td>
<td>Vonhof and Hobson (2001)</td>
<td>53.46</td>
<td>-112.81</td>
</tr>
<tr>
<td>32</td>
<td>Content Bridge, Red Deer River</td>
<td>Lausen (2006a)</td>
<td>52.31</td>
<td>-113.08</td>
</tr>
<tr>
<td>33</td>
<td>C.F.B. Wainright</td>
<td>C. Lausen and L. Burt, pers. comm. (2008 baseline survey by Westworth Associates)</td>
<td>52.75</td>
<td>-111.17</td>
</tr>
</tbody>
</table>

1 Provincial Museum of Alberta
List of Titles in This Series
(as of May 2009)

No. 1  Status of the Piping Plover (*Charadrius melodus*) in Alberta, by David R. C. Prescott. 19 pp. (1997)


No. 4  Status of the Ord’s Kangaroo Rat (*Dipodomys ordii*) in Alberta, by David L. Gummer. 16 pp. (1997)


No. 8  Status of the Peregrine Falcon (*Falco peregrinus anatum*) in Alberta, by Petra Rowell and David P. Stepnisky. 23 pp. (1997)

No. 9  Status of the Northern Leopard Frog (*Rana pipiens*) in Alberta, by Greg Wagner. 46 pp. (1997)


No. 15 Status of the Plains Hognose Snake (*Heterodon nasicus nasicus*) in Alberta, by Jonathan Wright and Andrew Didiuk. 26 pp. (1998)

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No. 35  Status of Soapweed (Yucca glauca) in Alberta, by Donna Hurlburt. 18 pp. (2001)

No. 36  Status of the Harlequin Duck (Histrionicus histrionicus) in Alberta, by Beth MacCallum. 38 pp. (2001)


